- Quantifying river-groundwater interactions of New
 Zealand's gravel-bed rivers: The Wairau Plain.
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Abstract

New Zealand's gravel rivers have deposited coarse, highly conductive gravel sedi-6 ments that are covered with relatively poorly developed thin soils. The shallow 7 groundwater in these stratified gravel aquifers is predominantly fed by river water. 8 Recharge mechanisms in these rivers are poorly understood and the management of g the groundwater resources is challenging, particularly under a more variable future 10 climate. To better understand the river-groundwater exchange processes in these 11 rivers, we investigate the Wairau Plain aquifer which is pumped for drinking water 12 and irrigation. A three-dimensional surface water - groundwater flow model (MOD-13 FLOW) has been set up using a revised geological model of the Wairau Plain. The 14 model was calibrated using targeted field observations, "soft" information from ex-15 perts of the local water authority, and the model-independent parameter estimation 16 software PEST. We determined the trade-off between data fit and parameter ho-17 mogeneity using regularization techniques and the PEST Pareto analysis tool. The 18 calibrated model performs well for both the calibration data set and independent 19 data. Flux-weighted transit-time distributions for the largest spring at the Wairau 20 Plain were calculated using particle tracking. Mean transit times of less than 1 yr 21 suggest very young water for Spring Creek. The uncertainty of the model simula-22 tions was evaluated using Null-space Monte-Carlo methods and is larger for mean 23 travel times than the for river-groundwater exchange flows. Our analysis suggests 24 that the river is hydraulically perched above the regional water table in its upper 25 reaches and is gaining downstream where marine sediments overlay the unconfined 26 gravels and where most of the springs of the Wairau Plain emerge. The net re-27 charge to the Wairau aquifer is on average 7.3 $m^3 s^{-1}$. Although the river discharge 28 is highly dynamic and regularly exceeds $1000 \text{ m}^3 \text{s}^{-1}$, the net exchange flow is capped 29 and rarely exceeds $12 \text{ m}^3 \text{s}^{-1}$. Changes in aquifer storage are mainly affected by the 30 frequency and duration of low-flow periods in the river. This study significantly im-31 proved our understanding of Wairau river-groundwater exchange mechanisms. We 32

hypothesise that the methodology and general mechanisms are transferable to other

³⁴ New Zealand rivers with similar characteristics.

35 1 Introduction

Many New Zealand rivers flow from mountain valleys onto alluvial plains where they have 36 deposited Quaternary gravel sediments of varying thickness (Rosen and White, 2001). 37 These rivers lose water to shallow, unconfined aquifers formed by the alluvial fans and 38 gain water near the coast as groundwater moves into confined aquifers and returns to 39 the surface (e.g. Larned et al., 2008). Lowland aquifers are often an important water 40 resource for municipal, agricultural, and industrial uses (e.g. Brown et al., 1999; Rosen 41 and White, 2001). The management and protection of these water resources requires a 42 good understanding of the interacting processes, particularly the quantification of river-43 groundwater exchange rates and their prediction under changing environmental conditions. 44 River recharge can be the major source for groundwater in gravel-bed river systems and 45 land-surface recharge is typically much lower. For the Heretaunga Plains aquifer, for 46 example, the annual rainfall recharge is only 3% of the river recharge (Dravid and Brown, 47 1997). Less than 20% of estimated Avon River base flow is rainfall recharge (White, 2009). 48 Although the research on surface-subsurface exchange processes has increased dramatically 49 towards the end of the last century (Stanley and Jones, 2000), the understanding and 50 quantification of the interaction processes present still a major challenge (Sophocleous, 51 2002; Brunner et al., 2011; Lamontagne et al., 2014). Field techniques to quantify river-52 groundwater exchange rates encompass, amongst others, the measurement of the hydraulic 53 gradient between the river and the adjacent groundwater, dilution tests with chemical or 54 heat tracers, pumping or slug tests, and mass balance approaches (for a comprehensive 55 review c.f., Kalbus et al., 2006; Rosenberry and LaBaugh, 2008; González-Pinzón et al., 56

2015). Some of these methods are rather elaborate and time-consuming and also difficult 57 to transfer to larger scales. Differential stream flow gauging is a more readily applied 58 mass balance method for the river-reach scale, where the net loss/gain over the length 59 of a river section is determined by the difference between gauged flows at an upstream 60 and a downstream cross-section. However, the flow channels of gravel-bed rivers in New 61 Zealand are typically braided which could require simultaneous flow measurements in 62 multiple braids (White et al., 2001). In addition, the mass balance approach requires 63 the quantification of all other sources and sinks along the river reach such as tributaries, 64 potential water takes, and underflow within the confines of the active river bed. 65

A controlling factor to determine the exchange rates between surface water and groundwa-66 ter is the state of connection between the two compartments (Brunner et al., 2009, 2011). 67 This is often poorly understood in the field. In addition, the state of connectivity might 68 change due to the more dynamic nature of river flows and a delayed reaction of ground-69 water levels. Therefore, Brunner et al. (2011) called for more field studies dealing with 70 the state of disconnection. In New Zealand, several studies have been dedicated to invest-71 igate river-groundwater connectivity and river-groundwater exchange flows (e.g., Brown 72 et al., 1999; Larned et al., 2008; Rupp et al., 2008; White, 2009; White et al., 2012; Close 73 et al., 2014). The Selwyn River is a prime example for highly complex spatio-temporal 74 flow patterns and various states of connectivity and large river water losses in the alluvial 75 plains (Larned et al., 2008; Rupp et al., 2008). Other New Zealand river systems showed 76 consistent flow patterns over larger periods of times. Differential discharge measurements 77 taken between 1957 and 1995 at a 3 km section of the Ngaruroro River show a consistent 78 loss of 4.3 $m^3 s^{-1}$ for river flows below 35 $m^3 s^{-1}$ (Dravid and Brown, 1997). Similarly, 79 consistent flow losses were reported for sections of the Rakaia River and the Waimea River 80 by White et al. (2001). 81



time/space scale, the estimation of river-groundwater exchange rates are often compli-83 mented by hydrological modelling. Numerical models can be used to integrate field obser-84 vations of various types and to investigate scenarios for (regional) water management (e.g., 85 Panday and Huyakorn, 2004; Kollet and Maxwell, 2006; Jones et al., 2008; Spanoudaki 86 et al., 2009; Maxwell et al., 2015). Several competing models and modelling schools have 87 been discussed in the scientific literature (e.g., LaBolle et al., 2003; Furman, 2008) which 88 is not repeated here. A comprehensive review of regional integrated models has recently 89 been presented by Barthel and Banzhaf (2016). Integrated models, that simulate both sat-90 urated and unsaturated flow, as well as surface water, groundwater and the full coupling 91 between them in a physical way (Brunner et al., 2010), can be highly accurate. Yet they 92 require a large amount of data for their parametrization and their practical application 93 is often restricted by large run-times (von Gunten et al., 2014). This is particularly chal-94 lenging in braided river systems where the flow channel geometry is extremely complex 95 and frequently changing over time. Some attempts were made to generate braided river 96 terrain models in New Zealand for the Rees River and the Waimakariri River using air-97 borne photography, LIDAR, and multi-point statistics (Pirot et al., 2014; Williams et al., 98 2014, 2016) but these methods are far from being routinely applied in surface water -99 groundwater modelling. 100

On the other hand, conceptual models that treat subsurface compartments as reservoirs are 101 less data-hungry, have less parameters and are typically much faster. In the New Zealand 102 context, Yang et al. (2017) introduced an additional conceptual groundwater store to 103 the national hydrological model TopNet (Bandaragoda et al., 2004; Clark et al., 2008) 104 to account for water transfer from rivers and also for cross/inter-catchment groundwater 105 flow. However, river losses/gains are inputs to the TopNet model and are considered to be 106 constant over time. This limits the potential application of TopNet to river basins where 107 the river-groundwater exchange flows are known and time-invariant. 108

The numerical model MODFLOW is most frequently used to simulate surface water -109 groundwater interactions (Furman, 2008). MODFLOW (Harbaugh, 2005) in its calcula-110 tions distinguishes between hydraulically connected and disconnected states and gener-111 ally constitutes a good compromise between fully coupled models and conceptual models. 112 Brunner et al. (2010) have revised the assumptions of MODFLOW in the context of sim-113 ulating surface-water - groundwater interactions and provided some guidance about its 114 application. In a later study it was concluded that the behaviour of disconnected river 115 can often be approximated by neglecting the unsaturated zone (Brunner et al., 2011). 116 MODFLOW has previously been applied in New Zealand (e.g. Fenemor, 1989; Baalousha, 117 2012; Gusyev et al., 2013). 118

¹¹⁹ In our study, we also use MODFLOW as a simulation tool to analyse and quantify sur-¹²⁰ face water - groundwater interaction in gravel-bed rivers. The aims of this study are ¹²¹ summarized as follows:

- to present modelling techniques for integration of hydrological data of various types with the specific focus on understanding river-groundwater exchange flows,
- to perform a detailed investigation of the spatial and temporal variation of the exchange flows, their dependence on river flows, and the state of connection between river and groundwater, and

to assess parametric and predictive uncertainty on simulated groundwater heads, net
 exchange flows, as well as spring flows and transit times using rigorous yet pragmatic
 methods suitable for highly-parametrized models.

We will demonstrate our approach for a section of the Wairau River on the Wairau Plain, which interacts strongly with the shallow, stratified and highly-conductive gravel aquifer. The aquifer is managed by the Marlborough District Council (MDC) and is of regional importance because it supplies all of the municipal water requirements for Blenheim, Renwick, and Woodbourne, together with most of the vineyard irrigation supply.

The reminder of the paper is structured as follows. First we present the study area and the corresponding MODFLOW model. Then we describe the calibration strategy including the various calibration targets and our proposed uncertainty quantification methodology. In Section 3 we present the results of the model calibration, the analysis of the rivergroundwater exchange mechanisms, as well as model predictions of transit-time distribution for the largest Spring on the Wairau Plain. The paper is concluded by a synthesis of our findings.

¹⁴² 2 Materials and Methods

¹⁴³ 2.1 Wairau Plain

The study site is located in the lower reaches of the Wairau River catchment in the 144 Marlborough District of the northern South Island, New Zealand. The Wairau River 145 basin drains an area of 3430 km^2 which is covered by a mix of exotic pine and native 146 beech forest in the northern and western ranges (elevation up to 2300 m) and pasture and 147 shrub-lands in the southern hills. Just prior to discharging into the Pacific sea, the Wairau 148 River enters the Wairau Plain which is New Zealand's largest wine growing area. Here, the 149 braided gravel-bed river flows since modern times in a 100 - 200 m wide floodway at the 150 northern edge of the Plain with constructed stop-banks as much as 1 km apart (Figure 1). 151 The elevation of the Plain ranges from 72 m.a.s.l. in the West to sea level in the East over 152 a distance of roughly 27 km. The river almost exclusively feeds the underlying Wairau 153 aquifer which serves as the major resource for drinking water and irrigation in the region. 154 It is the most extensive and important resource in the region by far and ranks amongst 155

the most significant aquifers in New Zealand (Davidson and Wilson, 2011). A slow, but constantly declining trend in aquifer levels and spring flows have been observed over the past decades, which has triggered this investigation aimed at a better understanding of the recharge mechanism and the river-groundwater interactions.

In this study, we focus on a 22 km long section of the Wairau River that encompasses the entire recharge area and the majority of the Wairau Plain. It covers the river reach from downstream of the Waihopai River confluence to the SH1 bridge upstream of the Tuamarina River confluence (Figure 1).

Geology The earliest investigations of the Wairau Plain geology were carried out by 164 Brown (1981). More recently, a detailed 3D geological model of the coastal Wairau Plain 165 geology and its deeper aquifer structure was presented by Raiber et al. (2012). The 166 basement geology of the Wairau basin consist of schist in the North and greywacke in 167 the South. These rocks are overlain by a sequence of Pliocene to Pleistocene glacial 168 outwash gravels interspersed with interglacial marine horizons at the coast. The youngest 169 of these gravels is the Speargrass Formation, which is considered to form the base of the 170 Wairau Aquifer. The Wairau Aquifer is hosted by high permeability Holocene sediments 171 of the Rapaura Formation. These gravels have been formed by alluvial reworking of the 172 Speargrass Formation and are orders of magnitude more transmissive. Towards the coast, 173 the Rapaura Formation is overlain by marine silts of the Dillons Point Formation, which 174 form a confining horizon to the Wairau Aquifer (Brown, 1981). More recently, Wilson 175 (2016) reviewed the geological records of the Rapaura Formation for a more detailed 176 analysis of its internal structure. Structure contours of the Speargrass Formation surface 177 indicated that the Rapaura Formation has a maximum thickness of 30 to 35 m, and is 178 typically 20 m over most of the aquifer. Three lithological members were distinguished, 179 with some lateral variability evident in the uppermost member: 180

- Upper Member: 8 ± 3m of mostly stratified gravels of moderate permeability incised locally by facies of high permeability associated with recent alluvial channels.
 Low Permeability Member: clay-rich gravels 3-9 m thick, deposited as over-bank flow deposits when sea levels began to stabilise about 6.5 ka.
- Lower Member: high permeability alluvial gravels 9.5 ± 5m thick deposited 9.5 to 7
 ka during a period of warming global temperatures.

Based on the identified stratigraphy, a new conceptual model for the internal structure of the part of the Rapaura Formation that underlies the study area was developed (Figure 2). The soils of the Wairau Plains are typically shallow, stony and well-draining. They can be classified in nine groups as depicted in Figure 1.

Hydrological data The flow record of the Wairau River close to the SH1 bridge dates 191 back several decades. In June 2014, MDC staff installed three additional temporary re-192 corder sites upstream of SH1, where the river flows in a single braid. These sites are 193 subsequently referred to as Rock Ferry, SH6, and Wratts Rd (Figure 1). River stage is 194 measured at these sites and then converted into discharge using elaborately established 195 rating curves. The flow ratings at these sites have been renewed after each larger flow event 196 because of changes in the braided river bed geometry. Spot gaugings of Wairau River flow 197 were conducted at these and other sites as far back as in the 1970s. The gaugings were 198 conducted usually under low-flow conditions and at the same day. The differential gauging 199 allows to examine river losses and gains. Both historic and recent gaugings show consist-200 ently the river losing water between Rock Ferry and Wratts Rd and then gaining between 201 Wratts Rd and SH1 (Figure 3). 202

At the intersection of the Rapaura and Dillons Point Formations, groundwater is forced to the surface and emerges as springs across the Wairau Plain. The major spring on the Plain is Spring Creek which has a mean flow of about 4.0 m³/s at the Motorcamp recorder site (Figure 3). The record of manual gaugings dates back to 1990 and was complemented in 2013 by an automatic recorder. However, only manual gaugings are used in this study since the automatic recorder had frequent malfunctions and was at least in some cases influenced by channel blocking and weeds.

Groundwater levels are observed at four permanent (3009, 3821, 3954, 4577) and six 210 temporary MDC wells (903, 7007, 1685, 1690, 1696, 10426) distributed over the length of 211 the Wairau Plains (Figure 3). The wells are screened at different depths and across all 212 three of the main facies of the Rapaura Formation. The temporary wells were equipped 213 in January 2016 specifically for this project. Data gaps occurred in wells 7007, 1685, and 214 10426 in December 2016 and January 2017. Some additional spot measurements (manual 215 dipping) were taken in wells 7007 and 10426 prior to January 2016. All permanent wells 216 are part of the MDC core monitoring program and have continuous long-term records. 217

Meteorological data The Wairau Plain receives on average 650 mm of annual precipitation. The mean annual temperature is 12.8 °C and the sun shines on average 6.7 hours per day. This unique climate makes the area so attractive for winegrowers. The meteorological data required for our calculations was sourced from the Blenheim Research Station.

All the continuous data was aggregated / averaged to daily values for use in our model simulations. To integrate the new information from the temporary logger sites, we have chosen the time period between 1/7/2013 and 20/02/2017 in our investigation.

226 2.2 Wairau Aquifer model

A transient surface water - groundwater model for the study area was set up in MODFLOW NWT which is designed to solve problems involving drying and re-wetting non-linearities

of unconfined groundwater-flow (Niswonger et al., 2011). The graphical user interface 229 ModelMuse (Winston, 2009) was used to set up the model domain and boundary condi-230 tions. A plan view of the model boundaries is shown in Figure 1. The total model area is 231 84.8 km². To the North, the domain is bounded by the northern bank of the Wairau River. 232 In the West, the domain starts at Rock Ferry, a natural rock constriction of the Wairau 233 valley. The southern boundary is normal to the regional groundwater level contours shown 234 in Davidson and Wilson (2011). The eastern boundary is drawn at the SH1 bridge, ap-235 proximately 5 km off the coast, because groundwater in the Rapaura Formation is forced 236 up through the confining Dillons Point Formation which forms a natural boundary. Deep 237 groundwater flow over the eastern face of the Rapaura Formation is considered constant 238 throughout time at a rate of - $0.7 \text{ m}^3 \text{s}^{-1}$ as estimated from spring flows in Grovetown 239 Lagoon to the East. The FHB package (Leake and Lilly, 1997) is used to implement the 240 corresponding boundary condition in the model. 241

The top elevation of the model domain is derived from a high resolution LIDAR image which was interpolated at the grid nodes of the MODFLOW computational grid. The bottom of the model domain is defined by the elevation of the Speargrass Formation (Figure 2) and is considered an impermeable zero-flux boundary. The northern, western, and southern boundaries of the model domain are considered no-flux boundaries too.

Figure 4 depicts the computational MODFLOW grid. As a result of a preliminary sens-247 itivity analysis with different grid sizes, we selected a regular cell size of 200×200 m 248 in our model. The geology was implemented by three computational layers matching the 249 formation boundaries. Thus the model domain consists of $3 \times 2120 = 6360$ active cells. 250 The first layer of the grid is considered the Upper member of the Rapaura Fm in the West 251 and the Dillons Point Fm in the East. On the surface, the intercept of the two formations 252 is marked as aquitard boundary in Figures 1 and 4. A minimum grid-cell thickness of 1.0 253 m was assumed for the cells of the intercept and also for all other grid cells for numerical 254

efficiency. According to the geological record, the Lower and Low permeability members of the Rapaura Formation outcrop underneath the Wairau River for a relatively short section in the West.

The Wairau River is implemented by the streamflow routing package SFR Wairau River 258 (Niswonger and Prudic, 2005) which can be used to simulate connected and disconnected 259 streams. Because of the highly permeable sediments of the gravel-bed river, we consider 260 head-dependent stream leakage when the river is connected with groundwater and unit-261 gradient flow when the river is disconnected. Further, we defined 12 different sections 262 along the river, with locations corresponding to locations of a detailed survey of the river 263 geometry conducted at 25 cross-sections between Rock Ferry and SH1. The SFR package 26 requires as input the time series of river discharge at Rock Ferry which was calculated 265 from the discharge record at SH1 plus a constant 7.64 $m^3 s^{-1}$ that was determined from 266 a correlation analysis using the concurrent record of stream flow ($R^2 = 0.98$). MODFLOW 267 calculates the actual river length for each river cell. The SFR package further requires the 268 parametrization of the stream-bed hydraulic conductivity, thickness of the river bed, and 269 two geometric functions describing the functional relationship between river stage $h[\mathbf{m}]$, 270 wetted perimeter of the stream channel L_{wp} [m²], and discharge in the river Q_{riv} [m³/s]: 271

$$h(Q_{riv}) = aQ_{riv}^b; \quad L_{wp}(Q_{riv}) = cQ_{riv}^d , \qquad (1)$$

where, *a*, *b*, *c*, *d* are empirical constants. These constants were initially derived for each of the 25 river cross-sections, simultaneously put into the model, and simulation results compared to a single set of parameters derived for an average cross-section and applied to all cross-sections in the model. The sensitivity to the more detailed representation of the river geometry was low, because of the parameter interactions to stream-bed conductivity and the underlying hydraulic conductivity field. This allowed us to reduce the parametrization

effort and we applied a single set of river geometry parameters at 12 river cross-sections 278 in all our simulations: a = 0.192, b = 0.16, c = 4.83, d = 0.239. The thickness of the 279 river-bed was assumed to be 1 m in all river sections. A detailed (and transient) para-280 metrization of the geometry functions in the model is possible but in fact not desirable 281 for practical reasons, because acquiring the channel-bed geometry information involves a 282 significant experimental effort at regular intervals in gravel-bed rivers (after each major 283 flood). Modelling approaches to simulate the transient evolution of river-bed morphology 284 (Pirot et al., 2014; Williams et al., 2014, 2016) are still far from being routinely used. 285

Springs and Streams The stream network and the springs emerging at the eastern 286 Plain are depicted in Figure 4. They are simulated using the DRN package (Harbaugh, 287 2005) which describes head-dependent flux boundaries. If the head in a drain cell falls 288 below a certain threshold, the flux from the drain to the model cell drops to zero. The DRN 289 package requires the specification of drain bed conductivity K_D and drain elevation. The 290 latter is derived from the LIDAR image and was offset by -1 m for the channel depth. Five 291 different sections of the springs and streams are distinguished (Figure 4): ND describes the 292 northern drain, a spring that discharges into the Wairau River just East of the aquitard 293 boundary. Spring Creek is divided at the flow gauging station in a western and an eastern 294 part (SC1 and SC2, respectively). Further, the Omaka River and the eastern Opawa River 295 at the southern model boundary (OR1) are distinguished from the western Opawa River 296 (OR2). Each drain section is parametrized separately. 297

Recharge and Irrigation Groundwater recharge from the land surface is considered as a specified flux boundary at the top of the model domain using the RCH package (Harbaugh, 2005). The land use of the Wairau Plain is predominantly vineyards which are irrigated using groundwater that is pumped locally from the aquifer. Irrigation abstraction is simulated using the WEL package which applies a specific flux boundary to internal ³⁰³ cells, here specifically cells in layer three. Groundwater recharge and irrigation demand
³⁰⁴ are computed for each of the nine soil types of the Wairau Plain using a soil water balance
³⁰⁵ model which is described in detail in Section 2.3.

³⁰⁶ 2.3 Land surface recharge model

The landuse at the Wairau Plain is almost exclusively vinyards. Land surface recharge and 307 irrigation demand are simulated using a daily soil moisture balance model, which has been 308 modified from the Rushton model (Rushton et al., 2006). The Rushton model is a simple 309 two-layer soil model which uses a near-surface soil store to enable evapotranspiration to 310 occur during soil moisture deficit conditions on days following rainfall events. Without 311 this near surface soil storage, evapotranspiration values following rainfall events would 312 be underestimated (de Silva and Rushton, 2007). The proportion of rainfall infiltration 313 that is partitioned to the near surface soil store is determined by an empirical coefficient, 314 f_s . Values of f_s are related to soil texture and drainage, and are zero for coarse sandy 315 soils, 0.4 for sandy loams, and 0.75 for clay loams (Rushton et al., 2006). For the soils 316 of the Wairau Plain we estimated values of f_s , ranging from 0.1 (gravelly sand) to 0.75 317 (deep clay loam). Soil texture and Total Available Water (TAW) values were sourced from 318 the New Zealand Fundamental Soil Layer database (Landcare Research, 2000). Readily 319 Available Water (RAW) for vineyard grapes was assumed a value of 45% of TAW, following 320 Allen et al. (1998). Soil moisture in the deeper soil layer is calculated after near-surface 321 evapotranspiration has been accounted for in the near surface soil store, S. The soil 322 moisture deficit, SMD for each day with index i is calculated for each soil type as follows: 323

$$SMD_i = SMD_{i-1} - \Delta_i + S_i + AE_i \tag{2}$$

where, Δ_i is the balance of daily inputs to the soil:

$$\Delta_i = P_i - R_i + S_{i-1} \,, \tag{3}$$

and P_i and R_i are precipitation and surface runoff, respectively. Daily rainfall and poten-325 tial evapotranspiration (PET) values were taken from the record for Blenheim Research 326 Station. PET is derived by the Penman-Monteith equation (Allen et al., 1998) as grass 327 reference evapotranspiration ET_0 . A seasonally-varying crop factor, K_c , was applied for 328 vineyard grapes based on sap flow measurements in a Marlborough vineyard (Green et al., 329 2014). Actual evapotranspiration (AE) is assumed to equal PET when soil water is readily 330 available. For RAW < SMD < TAW, the vineyard becomes water-stressed and transpires 331 at a reduced rate unless inputs to the soil exceed *PET*. This situation is represented in 332 the model by applying a water stress factor: 333

$$AE_i = K_{S,i}K_{c,i}ET_{0,i}, \qquad (4)$$

334 with

$$K_{s,i} = \frac{\text{TAW} - \text{SMD}_{i-1}}{\text{TAW} - \text{RAW}}.$$
(5)

If the soil moisture content reaches the value of TAW, the roots are unable to extract water, 335 and $AET = \Delta$. Drainage to groundwater occurs only in the model when SMD is negative, 336 i.e. when there is surplus water in the soil moisture reservoir. Soil moisture calculations 337 are started during winter conditions so that an initial soil moisture deficit of zero can be 338 assumed. This enables a lead-in time for the model to establish a suitable initial condition 339 for the beginning of the first calendar year. Surface runoff R is calculated by the SCS 340 method (Soil Conservation Service, 1972). It is assumed that 2.2 mm vineyard irrigation 341 occurs on days when soil moisture is less than 70% of RAW during the irrigation season 342

³⁴³ (October to April). This irrigation threshold was determined by comparing modelled ³⁴⁴ irrigation demand with water meter data from vineyards on the Wairau Plain.

345 2.4 Parametrization

The model domain and the boundary conditions described in Section 2.2 require various parameters to be specified. The corresponding parametrization scheme is described in this section.

The three lithological members of the Rapaura Formation have different hydraulic proper-349 ties which are considered in the model by an independent parametrization of the hydraulic 350 characteristics for the three layers. There is also considerable horizontal heterogeneity of 351 aquifer properties (Davidson and Wilson, 2011; Wilson and Wöhling, 2015) which is im-352 plemented in each layer using a pilot point parametrization technique (e.g., Doherty 2003; 353 Doherty et al. 2010) for the hydraulic conductivity and specific yield fields. Pilot points 354 are discrete, user-defined locations throughout the model domain that are used here for 355 cell-by-cell parametrization of the saturated hydraulic conductivity K_H , and of the specific 356 yield S_y through interpolation from the pilot points to the model grid. Corresponding to 357 only regional changes in horizontal heterogeneity we used a exponential variogram with 358 range 5 km and 26, 31, and 33 pilot points at a regular spacing for the Upper, Middle, and 359 Lower member of the Rapaura Formation, respectively. Given the much lower hydraulic 360 conductivity of the confining Dillons Point Formation compared to the Rapaura fm, uni-361 form properties are assumed for the confining layer. Further, a uniform anisotropy factor 362 for the hydraulic conductivity, f_a , and uniform specific storage, S_s , was assumed for each 363 of the four geological units. 364

Other parameters to be considered in the model are the vertical hydraulic conductivity for each of the 12 defined river sections, K_R , and the drain bed conductivity, K_D , of each ³⁶⁷ of the five drain sections as defined in Section 2.2.

³⁶⁸ 2.5 Model calibration and uncertainty analysis

In total, there are 207 parameters for the Wairau Plain model (Table 1). These parameters can't be measured directly at the required spatial and temporal scales and thus effective parameter values need to be estimated trough model calibration. For highly parametrized models like the one presented, automatic model calibration is the only feasible option. In this study we used the model independent parameter estimation software PEST (Doherty, 2016b,c) which is ideally suited for highly parametrized inversion problems (Doherty et al., 2010).

Calibration data The objective of the model calibration in general is to minimize 376 the discrepancy between model simulations and measured data. In our study we used 377 observations of groundwater head, Spring Creek flows, a spot measurement of differential 378 river flow gauging, and three "soft targets" which contain expert knowledge from MDC 379 groundwater scientists. The data set is separated in 123-day lead-in period, a 925-day 380 calibration period and a 284-day evaluation period (Table 2). Approximately 70% of 381 the head observations from the four permanent observation bores are used for model 382 calibration and the reminder (30%) for model evaluation. In contrast, the majority of 383 the head observations from the six temporary bores (between 59 and 72%) are used for 384 model evaluation. Another calibration target was formed on the basis of the historic 385 differential river gaugings (Figure 3). It follows the rationale that the average river losses 386 and gains between Rock Ferry and SH1 have been observed to be almost constant during 387 low flow periods and consecutive dates. A low flow period is present in the calibration 388 data set between 31/01/2014 and 15/03/2014 ($\bar{Q}_{riv} = 14.4 \,\mathrm{m^3 s^{-1}}$ at SH1). The mean 389 river exchange flux for that period and the river section between Rock Ferry and Wratts 390

Rd (losing section), $Q_{ex,13}$, is assumed to correspond to the mean loss from the historic 391 measurements which is $5.73 \text{ m}^3/\text{s}$. This constitutes the first soft target in our model 392 calibration. Secondly, the mean flows in the river reach between Wratts Rd and SH1, 393 $Q_{ex,3}$, is targeted at a net gain of -0.5 m³/s. Please note that flows out of the model 394 domain are negative numbers and fluxes into the model domain are positive numbers. 395 Other calibration targets specified from expert knowledge are a mean gain of Spring Creek 396 flows downstream of the gauging station at Motorcamp of $Q_{SC2} = -0.5 \,\mathrm{m^3 s^{-1}}$ and a mean 397 gain of all the southern streams of $Q_{SS} = -1.5 \,\mathrm{m}^3 \mathrm{s}^{-1}$. 398

Objective function PEST uses a sum-squared error (SSE) objective function, that 399 can be weighted by the measurement error (Doherty, 2016b). Using different physical 400 quantities (data types) with different numerical ranges and different observation numbers 401 in a SSE objective function leads typically to unequal weighting of the different data 402 types. The weighting of the individual observations is therefore of great importance for the 403 outcome of the calibration. Weighting of data expresses the degree of belief the modeller 404 has in the individual pieces of information and is therefore to some extent subjective. The 405 weighting of the different data types was determined by trial-and-error to obtain a balance 406 between "hard" and "soft" calibration targets. The weights of individual data points are 407 reported in Table 2 and used in all our model calibration runs. 408

Parameter regularization Regularization techniques are used in order to constrain potential solutions of the model calibration and to avoid unrealistic artefacts in spatially correlated data (e.g., Doherty 2003; Moore 2005). By regularization, parameter fields are penalized when deviating from the spatial correlation defined, for example, by a variogram. We applied Tikhonov regularization to the K_H and S_Y fields as well as to K_R using the variogram described in Section 2.4. The smoothness of a parameter field can be expressed by a weighted sum of parameter differences at neighbouring pilot points with weighting factors determined by the variogram. Deviations from "smoothness" is measured by a penalty objective function that has an optimum value of zero for a homogeneous field. The PEST groundwater utilities PPK2FAC, FAC2REAL (Doherty, 2016a), and the PEST utility ADDREG1 (Doherty, 2016c) are used to calculate the weighting factors for the pilot point locations and to implement a corresponding regularization objective function into the parameter estimation process with PEST.

Pareto optimization For reasons not further discussed here, there is typically a trade-422 off between the model's ability to correctly reproduce the data of local measurements and 423 the smoothness of parameter fields. The aim of the regularized parameter inversion (i.e. 424 the model calibration technique used here) is to find a compromise between data and reg-425 ularization objective functions and thus avoid overfitting. For example, it is not desirable 426 to place too much confidence in data that might not be represented in the model (e.g. by 427 subscale effects). On the other hand, we want to include as much spatial heterogeneity 428 in the calibrated model, as is legitimately supported by the data. The ideal weighting 429 between data and regularization objective functions is difficult to define. It could well be 430 argued that ideal weighting does not exist because the choice always involves some degree 431 of subjectivity by the modeller. To guide the choice, the trade-off between the object-432 ive functions can be determined using multiobjective calibration methods (e.g., Wöhling 433 et al., 2008; Reed et al., 2013) which result in a set of Pareto efficient solutions. These 434 solutions have the property that moving from one to another along the tradeoff surface 435 results in the improvement of one objective while causing deterioration in at least one 436 other objective (see Gupta et al., 1998 and others for further information on the Pareto 437 optimality). The Pareto optimization concept was adapted for highly parametrized inver-438 sion (Moore et al., 2010) and implemented in PEST. The method is used in this study to 439 simultaneously calibrate the model parameters and calculate the trade-off between data 440

and regularization objective functions described above. Subspace projection techniques
to increase the computational efficiency of the highly-parametrized model inversion were
also trialled here (SVD-Assist, Doherty et al., 2010; Doherty, 2016c), but the combination
of techniques lead to parameters being frozen at their boundaries, which was an undesirable effect. All our calibration runs were conducted using the parallel computing tool
BEOPEST. The parameter ranges were derived from geological information and expert
knowledge (Table 1).

The result of the calibration is a Pareto efficient set of solutions which was filtered to meaningful trade-offs by the concept of ϵ -dominance (Kollat et al., 2012). Finally, a compromise solution that exhibits both a good data fit and realistic parameter fields was subjectively selected from the Pareto set which then constitutes the calibrated model.

Uncertainty analysis After model calibration, an uncertainty analysis was performed 452 to assess the robustness of the model calibration and the reliability of model simulations 453 and predictions. Highly parametrized model calibration rarely leads to unique parameter 454 estimates, because of the insensitivity of model outputs corresponding to historical obser-455 vations of system state to some parameters, excessive correlation with other parameters, 456 or both (Doherty and Hunt, 2009). Conceptually, the parameter space can be divided into 457 two subspaces, the solution space and the null space. The solution space comprises para-458 meter combinations that are informed by the available data set. The null space comprises 459 parameter combinations that have little effect on model outputs when superimposed on 460 the calibration parameter set (Moore and Doherty, 2005; Doherty and Hunt, 2009). Note, 461 however, that these parameter combinations may have an effect on model outputs that are 462 not contained in the calibration data set. Sampling the parameter null space and analys-463 ing the resulting model simulations is therefore an effective means to determine non-linear 464 predictive uncertainty and is superior to linear first-order second moment (FOSM) meth-465

ods. Null space Monte-Carlo (NSMC) sampling utilities and FOSM predictive uncertainty
estimation utilities are readily implemented in PEST (Tonkin and Doherty, 2009; Doherty et al., 2010; Doherty, 2016c). NSMC sampling is applied in this study to estimate
post-calibration predictive uncertainty.

Model predictions Two types of model predictions are distinguished in this study. 470 The first type consists of data types that are already contained in the data set (here: 471 groundwater heads and Spring Creek flows) where predictions are made for different times 472 and different model forcings. These predictions are subsequently referred to as type I 473 predictions. The root mean squared error (RMSE) and the coefficient of determination 474 (\mathbf{R}^2) were used as metrics to summarize the model performance for these prediction types. 475 The second prediction type comprises model predictions / data types that are not fully 476 contained in the calibration data set. These predictions are subsequently referred to as 477 type II predictions. In this study, we predict the transient net-exchange flows between 478 Rock Ferry and SH1 (only low-flow means were used as soft-target in the calibration) 479 as well as the transit time distribution and mean transit time for Spring Creek water 480 upstream of the flow gauge. 481

Transit time distributions were calculated using reverse particle tracking methods with MODPATH (Pollock, 2012). The resulting particle tracks and residence times were postprocessed to calculate cumulative flux-weighted transit time distributions:

$$cdf_{TT} = \frac{1}{Q_T} \int_{i=1}^{N_p} \tau_i \cdot q_i , \qquad (6)$$

where, $i = 1 \dots N_p$ denotes the particle index, N_p is the total number of particles, τ_i is the particle travel time, q_i is the flux in the cell where the particle originates, and Q_T is the total flux, i.e. the sum of all q_i . The flux-weighted mean transit time (MTT) is then 488 calculated as the 50% quantile of the cdf_{TT} .

489 **3** Results and Discussions

First up in this section, the performance of the calibrated model and the uncertainty of type I predictions is analysed. In the second sub-section, we discuss parameter uncertainty and the plausibility of calibrated parameter values. Then, the river-groundwater exchange mechanisms for the considered section of the Wairau River (type II prediction) are analysed. Finally, we present the results of the other type II prediction, namely the transit time distribution and mean travel time of Spring Creek flows.

⁴⁹⁶ 3.1 Model calibration and evaluation

⁴⁹⁷ 3.1.1 Trade-off between data and regularization objective functions

The model was calibrated using a data objective function (OF_{dat}) and a regularization 498 objective function (OF_{reg}) . Figure 5 shows the trade-off between the two objective func-499 tions. Open circles depict all Pareto solutions obtained by the model calibration, while the 500 orange solutions depict the ϵ -dominant solutions which were used for the analysis. One 501 purpose of ϵ -dominance is to truncate meaningless solutions at the ends of a Pareto front, 502 where a small change in one objective function leads to a large change in at least one other 503 objective function. This is the case in here along the x-axis in Figure 5, where a small 504 improvement in the data fit leads to a rather large distortion of the parameter fields as 505 penalized by the regularization objective function. 506

⁵⁰⁷ Overall, we observed a large trade-off between data and regularization objective functions ⁵⁰⁸ which is demonstrated by the rather curved shape of the Pareto front. A more angular ⁵⁰⁹ shape of the Pareto front would indicate less trade-off between the two. The visual in-

spection of the parameter fields of the ϵ -dominant solutions revealed strongly distorted 510 fields for $OF_{reg} > 600$ that are a strong indication of over-fitting (results not shown). On 511 the other hand, parameter fields became unrealistically smooth for $OF_{reg} < 200$ while the 512 data fit deteriorated quickly. Therefore, we subjectively selected a compromise between 513 the two objective functions ($OF_{dat} < 60.6$, $OF_{reg} = 378.3$, indicated in blue in Figure 5). 514 Note, that the compromise solution can be selected in an objective manner by determining 515 the Pareto solution with the least Euclidean distance to the origin (for more details see 516 e.g., Wöhling et al., 2013). The parameter set of the compromise solution is subsequently 517 referred to as the calibrated model and used subsequently for analysing model perform-518 ance. NSMC simulations were conducted with that solution as described in Section 2.5 519 to access predictive uncertainty. The performance of the calibrated model is reported in 520 the next section while the parameter set and the corresponding uncertainty is discussed 521 in Section 3.1.3. 522

⁵²³ 3.1.2 Model performance

Groundwater heads Simulated groundwater heads obtained with the calibrated model 524 are compared to observations at the permanent and temporal MDC wells. Results are 525 summarized in Figures 6 and 7 and in Table 3 and are subsequently described. The 526 groundwater wells are presented by location from West to East in the figures, following the 527 gradients of the land surface and the groundwater table. Model simulations and observed 528 groundwater heads are indicated by the blue and orange lines (dots), respectively. The 95%529 uncertainty bounds determined from the NSMC simulations are shaded grey. A vertical 530 dashed line indicates the divider between the calibration period (left) and the evaluation 531 period (right). To facilitate an better comparison of the temporal dynamics between wells, 532 a constant y-axis spacing of 6 m was used in all figure panels. 533

⁵³⁴ Overall, the calibrated model represents the regional groundwater surface well. There is a

gradient between approximately 57 m.a.s.l. in the West at well 903 (Figure 6) and 7 m.a.s.l. 535 in well 3954 in the East (Figure 7) which is well reproduced by the model simulations. 536 The temporal variability of the groundwater heads as well as the depth to the water table 537 generally decreases from West to East. The variability is largest in wells located close to 538 the river (wells 903, 1690, 1696, 7007) and lowest in wells that are located underneath 539 the confining layer (wells 4577 & 3954). The detail of the observed groundwater head 540 variability is reproduced satisfactorily for most wells. However, some discrepancies remain 541 in wells 903 and 1690 (Figure 6) which can be explained by a relatively short calibration 542 data record for these wells and by model structural uncertainty at the western boundary. 543 The structural uncertainty includes a (too) narrow model domain with surrounding no-544 flow boundaries in the East, potential groundwater inflow from the Waihopai River, and/or 545 an influence from Gibsons Creek (Figure 1). Correspondingly, the model-to-measurement 546 misfit is larger for these wells compared to the other wells (Table 3). 547

Taking the perspective of a regional analysis, the performance of the model is considered 548 satisfactory for these other wells, which is confirmed by low RMSE values (ranging between 549 0.05 and 0.31 m) and large R^2 values (ranging between 0.68 and 0.91) for the calibration 550 period (Table 3). The model performance during the evaluation period is similar, but 551 shows a slightly larger variability with RMSE values ranging between 0.05 and 0.49 m and 552 \mathbb{R}^2 values ranging between 0.66 and 0.91. The 95% uncertainty bounds generally cover 553 the observations except for the wells at the western boundary where simulated heads 554 are generally overestimated (biased) and exhibit the largest model-to-measurement misfit. 555 The uncertainty tends to increase with the temporal variability of the groundwater heads 556 and is lower for the wells under the confining Dillons Point Formation. 557

⁵⁵⁸ **Spring Creek** The largest spring on the Wairau Plain is Spring Creek with a mean flow ⁵⁵⁹ of about 4 m^3s^{-1} at the Motorcamp gauging station. Spring Creek is fed by upwelling

groundwater and originates at the interface between the highly conductive Upper member 560 of the Rapaura Formation and the confining Dillons Point Formation (Figure 4). The 561 relatively large variability of the flow record and the correspondence to the Wairau River 562 flows shown in Figure 8 suggest the existence of rapid subsurface flow paths. These 563 are not uncommon for New Zealand's gravel-bed rivers which form highly transmissive 564 networks called open-framework gravels (Dann et al., 2009). Recent field work in the 565 Wairau floodway supports the existence of open-framework gravels in the Upper member 566 of the Rapaura Formation. 567

The model simulations match the observed Spring Creek flows well, although the variability of the flows seems to be overestimated when evaluated by the manual spot gaugings (Figure 8). Data taken by a continuous stage recorder installed in January 2013, however, showed very fast responses of Spring Creek flows to Wairau River floods and that the variability of simulated spring flows could be realistic. The recorder data were not used in the model calibration, though, because of continuing experimental challenges (e.g., weeds) that lead to drift and bias in the flow record.

The RMSE values of simulated Spring Creek flows are 0.22 and 0.32 $m^3 s^{-1}$ for the calib-575 ration and evaluation period, respectively. It should be noted that only seven data points 576 were available in the evaluation period (Table 3). The 95% uncertainty bounds are too nar-577 row to capture all the observations which is potentially a result of the chosen uncertainty 578 quantification method. Following the NSMC procedure presented by Tonkin and Doherty 579 (2009), we applied a re-calibration step for the underlying parameters which might in this 580 case lead to an overly optimistic contraction towards the calibrated model parameters. 581 In addition, the sample of 100 NSMC simulations might be simply too small. On the 582 other hand, alternative uncertainty quantification methods based on stochastic parameter 583 sampling techniques are too time-consuming for application to highly-parametrized models 584 and therefore not further investigated here. 585

Soft Targets The fitness of the calibrated model to the soft targets is summarized in 586 Figure 9. The box plot shows the 50/95% uncertainty bounds by the boxes and whiskers, 587 respectively. Also shown are the target values in blue and the simulation of the calibrated 588 values in orange. A very good agreement between targeted expert knowledge and the 589 model and narrow uncertainty ranges are obtained for the flow in the downstream branch 590 of Spring Creek, Q_{SC2} , as well as for the average river-groundwater exchange flows under 591 low-flow conditions, $Q_{ex,1}$ and $Q_{ex,13}$. The results suggests that the model reproduces 592 both the upwelling of groundwater through the confining Dillons Point Formation and the 593 behaviour observed in the historic differential flow gaugings (Figure 3). 594

The flow target for the southern streams, Q_{SS} , is overestimated by the calibrated model 595 and has larger uncertainty bounds. The target was based on historic stream gaugings in 596 the ephemeral Opawa River prior to a diversion scheme into Gibsons Creek became oper-597 ational. Smaller springs and drains that exist South of Spring Creek are not considered 598 in the model, which would in part explain the discrepancies together with structural un-599 certainties of the southern no-flow boundary. However, the focus of the study is on the 600 river-groundwater exchange fluxes and the soft targets are weighted less compared to other 601 types of data in accordance to the subjective belief (or its counterpart uncertainty) of the 602 information (Table 2). We have found that the inclusion of expert knowledge in our 603 model calibration is highly valuable for both constraining the parameter space, and for 604 establishing a degree of trust in the calibrated model. 605

606 3.1.3 Parameter uncertainty

The uncertainty of type I predictions was presented in the previous sections along with the performance of the model for the calibration and evaluation data set. The underlying parameter uncertainty of the calibrated model is presented in this section and is summarized in Figures 10, 11 and 12.

Hydraulic conductivity fields The left column of Figure 10 depicts the hydraulic 611 conductivity fields of the three members of the Rapaura Formation. It is reiterated here 612 that hydraulic conductivity (and specific yield) is only estimated at pilot point locations 613 which are then used for interpolation onto the MODFLOW grid which is presented in 614 the corresponding figures. Also shown are the Wairau River and the considered stream 615 network for orientation and the groundwater observation wells in the respective facies. 616 Consistent with the geological expertise, the Upper member of the Rapaura Formation 617 exhibits the largest K_H values while the low permeability member in the middle has a 618 somewhat lower permeability. Hydraulic conductivity seems to increase underneath the 619 Wairau River from West to East with a high-conductive zone downstream of Giffords Rd 620 connecting the river, Wratts Rd well and the Spring Creek area. The prediction of high-621 conductive zones in the Lower member of the Rapaura Formation is not easily understood, 622 but the overall pattern is consistent with earlier investigations of transmissivities derived 623 from well specific capacity by Davidson and Wilson (2011). The hydraulic conductivity of 624 the confining Dillons Point Formation is about three orders of magnitude smaller than the 625 maximum values in the Rapaura Formation, which is consistent with geological knowledge 626 and exploration results. 627

The uncertainty of the hydraulic conductivity fields is presented as one standard deviation 628 of K_H of the NSMC runs in the right panels of Figure 10. The uncertainty is relatively 629 large for some areas of the Upper and Lower members of the Rapaura Formation. It 630 is interesting to note that this does not result in an equally large uncertainty for type 631 I model predictions as was shown in the previous subsections. In some areas, the large 632 uncertainty is likely to be caused by insensitivity to model outputs (e.g., in the eastern 633 part of the Lower member). In other areas it may be caused by trade-offs in the fit to 634 different pieces of information in the calibration data set (in the Upper member). Since 635 the absolute value of K_H for the Dillons Point Formation is orders of magnitude smaller, 636

the uncertainty appears to be zero in Figure 10. This is not the case as shown below by
normalized parameter ranges.

Specific yield fields Figure 11 shows the specific yield fields of the calibrated model (left 639 panels) and their respective uncertainty (right panels). The S_y values are within expected 640 ranges for the coarse gravel materials of the Rapaura aquifer. Only little variability can be 641 seen in the parameter fields with two distinctive exceptions in the Northeast of the Upper 642 member and the West of the low permeability member. However, the uncertainty of the 643 S_y -fields is relatively large and uniform in all three members of the Rapaura Formation 644 with one standard deviation exceeding $\frac{1}{3}$ of the entire parameter range. This is also the 645 case for the Lower member, where the parameter values remained close to their starting 646 values. This suggests that the sensitivity of specific yield to the model outputs set is 647 relatively low in light of the calibration data set and that there is potential for parameter 648 simplification. 649

Other parameters The uncertainty of parameters that are not spatially correlated over 650 the entire model domain are depicted by box plots in Figure 12. Note that the parameters 651 are normalized by their respective ranges which are listed for convenience at the top of 652 the graph. The boxes and whiskers show again the 50% and 95% uncertainty bounds. 653 respectively. Also shown are the median (red lines) and the parameter values of the 654 calibrated model (blue dots). In some cases, the parameter values of the calibrated model 655 fall on their upper or lower boundary (Figure 12). These bounds represent meaningful 656 physical limits even though parameters are effective parameter values for the grid-cell 657 scale of 200×200 m. Although a better data fit would be possible, we didn't want to 658 increase the parameter ranges or introduce more fine-scale detail to the model, mainly 659 because we wanted to avoid overfitting. Some peculiarities of the calibration parameter 660 set are subsequently discussed. 661

The effective (uniform but unisotropic) hydraulic conductivity of the Dillons Point Form-662 ation is with $K_{aq} = 11.7 \,\mathrm{m^3 s^{-1}}$ about two orders of magnitude smaller than the average 663 in the Rapaura Formation. The value seems to be relatively high for the fine-textured 664 marine sediments. However, it should be noted that K_{aq} is an effective value that ac-665 counts for both flow through the pore matrix and flow along faster vertical passageways 666 for upwelling groundwater through the sediments. The existence of these pathways causes 667 the springs on the Wairau Plain to still gain water along their course to the East. Cor-668 respondingly, the effective value for the specific yield of the marine sediments is relatively 669 large $(S_{y,aq} = 1E^{-3})$ but the 95% uncertainty bounds for both K_{aq} and $S_{y,aq}$ cover almost 670 the entire range of expected values. 671

The specific storage for the three members of the Rapaura Formation S_{S1-3} is small and insensitive because the unit hosts unconfined groundwater. These parameters can be omitted from the model calibration - unlike the corresponding parameter for the confining Dillons Point Formation (S_{S4}).

The Upper and Lower members of the Rapaura Formation exhibit no significant difference in vertical vs. horizontal hydraulic conductivity ($F_{xz1} = 1.1, F_{xz3} = 1.4$). This is somewhat contradictory to data from bore logs and aquifer tests and suggests that groundwater head data perhaps isn't well suited to constrain anisotropy in unconfined sediments. In contrast, the Low Permeability member has a factor $F_{xz2} = 4.4$ lower vertical hydraulic conductivity (Figure 12) which corresponds well to layers of finer material interbedded in this unit as described in Section 2.1.

A regularization constraint was placed onto the spatial variability of river-bed hydraulic conductivity of the 12 river sections $(K_{b,R1} \dots K_{b,R12})$ to avoid overfitting and to make the model more robust to predictive bias. The calibration resulted in a deviation from the optimal regularization constraint, i.e. from all river-bed conductivities having the same value. In other words, the data has forced the pattern of the river bed conductivities which has a direct impact on river-groundwater exchange rates in the different river sections. In general, the river-bed conductivities increase from West to East (Figure 12) and are largest in the Wratts Rd area that coincides with the high-conductive zone in the Upper member of the Rapaura Formation described above. Together, these features form a highly transmissive passage of Wairau River water to Spring Creek.

⁶⁹³ 3.2 River-groundwater exchange mechanisms

The results of the previous section demonstrated that the calibrated model performs well to historic data (both for calibration and independent data sets) and that the obtained parameter set is in agreement with expectations, previous data and expert knowledge. This is a prerequisite for a trustworthy model in general and specifically if type II predictions are to be made by the model. The river-groundwater exchange flows are such a prediction. Results are summarized in Figures 13, 14, and 15 and are subsequently discussed.

700 3.2.1 Net exchange flows

Daily values of net river-groundwater exchange flows for the Wairau River section between 701 Rock Ferry and SH1, Q_{ex} , are presented in Figure 13b). The top panel depicts the corres-702 ponding Wairau River flows during the considered simulation period. The net exchange 703 flow is always positive and most of the time $Q_{ex} > 5 \,\mathrm{m}^3 \mathrm{s}^{-1}$ which means that overall, 704 the Wairau River is always losing water to the aquifer. Figure 13b) also shows that 705 the exchange flow is highly dynamic and correlated with the river flow. Large flood 706 events in the Wairau River typically also result in peaks for the exchange flow. However, 707 smaller flood events of less than $250 \,\mathrm{m^3 s^{-1}}$ at the end of prolonged low-flow periods in 708 summer also result in strong recharge peaks. One example is the relatively small flood 709 event $(Q_{riv} = 261 \,\mathrm{m^3 s^{-1}})$ on 17/03/2014 which occurred after a 7-week recession period 710

without any floods and caused a relatively large recharge peak of $16.2 \,\mathrm{m^3 s^{-1}}$. The much 711 larger river flood peak one month later $(18/04/2014, Q_{riv} = 967 \,\mathrm{m^3 s^{-1}})$ resulted in a re-712 charge peak that was similar in size $(Q_{ex} = 18.0 \,\mathrm{m^3 s^{-1}})$ compared to the previous event. 713 Similar examples can also be found in summer 2015 (08/03/2015, $Q_{riv} = 274 \,\mathrm{m^3 s^{-1}}$, 714 $Q_{ex} = 16.2 \,\mathrm{m^3 s^{-1}}$) and autumn 2016 (13/05/2016, $Q_{riv} = 330 \,\mathrm{m^3 s^{-1}}$, $Q_{ex} = 14.5 \,\mathrm{m^3 s^{-1}}$). 715 It is interesting to note that the relatively large parametric uncertainty (see previous Sec-716 tion) has only little effect on the predictive uncertainty of the net exchange flow. The 95%717 uncertainty bounds are very narrow and hardly discernible in Figure 13b). 718

Recharge flows greater than the $5 \,\mathrm{m}^3 \mathrm{s}^{-1}$ base line seem to be triggered already by even 719 smaller flood events and since they occur more frequent in winter and less frequent in 720 summer, the aquifer is mainly recharged in the winter months. To analyse this further, we 721 have depicted the exchange flow anomaly in Figure 13c). The anomaly is calculated as the 722 deviation of the cumulative net exchange flow from its mean during the simulation period. 723 Negative/positive gradients in the anomaly curve indicate exchange fluxes below/above 724 the mean, respectively. The seasonality is clearly visible in this representation of model 725 results. During summer, the gradient is negative indicating lower than average recharge. 726 During April - September (autumn/winter in the southern hemisphere) the gradient is 727 reversed indicating higher recharge and that the aquifer storage is re-filled during that time. 728 Groundwater heads are responding accordingly and show the same seasonality (Figures 6 729 and 7). 730

If the seasonal pattern of groundwater recharge from the river would be equal for each consecutive year, the anomaly curve would exhibit the same maximum and minimum value in each year. This is apparently not the case as seen in Figure 13c). There is inter-annual variability of rainfall in the Wairau catchment and thus also of aquifer recharge. 2014 and 2015 were particularly dry years on record which causes the seasonal maximum of the recharge flow anomaly to decrease for these years. However, the summer 2016 brought two major flood events in an usually dry period and was followed by a particularly wet winter and spring, which caused above-average aquifer recharge. It can be concluded from the analysis that time periods with frequent, consecutive river floods with return periods in the order of only weeks lead to enhanced aquifer recharge while prolonged dry periods cause lower aquifer recharge.

⁷⁴² 3.2.2 Spatial variability of river-groundwater exchange flows

To study the spatial variability of river-groundwater exchange flows along the Wairau 743 River, a snapshot of the model simulations was taken on 17/02/2014 which relates to 744 a low-flow period and a date where differential gaugings were conducted in the river. 745 Figure 14a) shows the groundwater head contours for that particular day. They are mainly 746 oriented from West to East following the gradient of the land surface. Some groundwater 747 mounding can be seen under the river between Rock Ferry, SH6, and half-way through 748 to Wratts Rd, which corresponds to a less transmissive area in the Upper Member of the 749 Rapaura Formation (Figure 10) and lower river bed conductivities in the upstream region 750 (Figure 12). 751

The simulated exchange flows for all river and drain cells in the model domain are depicted 752 in Figure 14b). Yellow and green colours indicate losses while blue colours indicate gains. 753 The analysis revealed that the largest river losses are to be found in an area half-way 754 between SH6 and Wratts Rd where the Upper member of the Rapaura Formation is 755 relatively thick. East of the line of confinement formed by the Dillons Point Formation, 756 all rivers and streams are gaining. Particularly high fluxes are visible at the origin of 757 Spring Creek and the lower reach of the Opawa River. To analyse the spatial pattern 758 further, we plotted in Figure 14c) the river-groundwater exchange flows along the path 759 of the Wairau River. The length and the up/downward direction of the bars indicates 760 the flow rate and losing/gaining conditions, respectively. Also shown are the river stage 761

and the groundwater table underneath the river. The picture confirms that the river is 762 losing in the first 18 km of the modelled section and that it is gaining downstream of the 763 location where the Dillons Point Formation is outcropping at the surface. The analysis 764 revealed further, that the river appears to be hydraulically disconnected (perched) over 765 long distances in the losing section. Head observations are mainly not located close to 766 the river (Figure 14a), but the projection onto the groundwater table underneath the river 767 shows a good agreement also for the confined area in the East, where the river is connected 768 to groundwater (Figure 14c). The largest river losses were predicted for the section between 769 SH6 and Wratts Rd $(Q_{ex2} = 4.77 \,\mathrm{m^3 s^{-1}})$ while the losses in the upstream section are lower 770 $(Q_{ex1} = 1.20 \,\mathrm{m^3 s^{-1}})$. Downstream of Wratts Rd, the river is at first still losing and then 771 gaining all the way to SH1, which results in a net gain of $Q_{ex3} = -0.54 \,\mathrm{m^3 s^{-1}}$. These 772 values are consistent with the differential gauging data taken on that day (Figure 3). 773

774 3.2.3 Correlation with river flow

Important information for resource management purposes is the functional relationship 775 between Wairau River flows and the net river-groundwater exchange, $Q_{ex,13}$. In order to 776 cover a larger range of hydrological situations, we have extended the model forcings of 777 the calibrated model and performed a forward simulation for the time period 1/1/2000778 to 20/2/2017. For each day in this simulation, $Q_{ex,13}$ is plotted over the corresponding 779 river discharge at SH1 in Figure 15. For clarity, the graph is truncated at Wairau river 780 flows of $120 \,\mathrm{m^3 s^{-1}}$. There seems to be a cap on the net exchange flows which don't exceed 781 $12 \,\mathrm{m^3 s^{-1}}$ for the data depicted here and rarely exceed $15 \,\mathrm{m^3 s^{-1}}$ even for the larger flows 782 in the simulation period (not shown). On the other hand, the net exchange flows are 783 relatively stable above a $5 \,\mathrm{m^3 s^{-1}}$ threshold for river flows greater than $20 \,\mathrm{m^3 s^{-1}}$. However, 784 when the river discharge at SH1 falls below $20 \,\mathrm{m^3 s^{-1}}$, a steep decrease of the net exchange 785 flows can be observed. This is an interesting result because the exchange flows seem to 786

vary throughout the year within a relatively narrow range of $5 - 8 \text{ m}^3 \text{s}^{-1}$ but are markedly decreasing during low flow periods (Figure 15).

River recharge is the major source of water for the Wairau Plain aquifer. In the considered 789 3.5-year simulation period, the land surface recharge was only 1% of the total water balance 790 and exhibited a strong seasonality (no recharge in summer). In wetter years, this value 791 might be slightly larger. However, the impact on aquifer storage seems to depend almost 792 exclusively on the river exchange flows, in particular on the frequency and duration of low-793 flow periods. A single large flood event doesn't counterbalance the net storage decrease 794 of extended dry periods which is also supported by the analysis presented in the previous 795 section (Figure 13). This means that extended dry periods could lead to a net aquifer 796 storage decrease which would require an above-average wet period with frequent, but not 797 necessarily large river floods to refill. 798

⁷⁹⁹ 3.3 MTT predictions to Spring Creek

More than half of the estimated mean river exchange flow of $7.3 \,\mathrm{m^3 s^{-1}}$ re-emerges in 800 Spring Creek. The spring is of great value for the community of the city of Blenheim for 801 recreational activities and its discharge and water quality is an important indicator for the 802 state of the shallow Rapaura aquifer. The age of the water is a supplementary measure 803 for estimating the risks and negative impacts associated with hydrological extremes (such 804 as droughts), catastrophic events (e.g., contaminant spills), and changes in land-use and 805 climate. The simulated flux-weighted transit-time distribution of the calibrated model is 806 depicted in Figure 16. Most of the water in Spring Creek appears to be older than 190 807 days. The mean transit time evaluated at the 50%-quantile of the cumulative density 808 function (cdf) is less than a year (MTT = 344 d). A distinctive tailing of the cdf suggests 809 a small contribution of water being older than two years. 810

The comparatively young age of the Spring Creek water is caused by the highly trans-811 missive subsurface zone between the Wairau River and the Spring Creek area as described 812 in Section 2.4 and depicted in Figure 10. Qualitatively, the MTT is in good agreement 813 with the analysis of previous and current water chemistry and isotope data (e.g., Davidson 814 and Wilson, 2011). The uncertainty of this type II model prediction is relatively large as 815 shown by the spread of the transit-time cdfs from the NSMC runs (grey lines in Figure 16). 816 The MTTs range between 222 and 421 days (histogram) and the mean of the MTT cdf is 817 at 315 days lower that the MTT of the calibrated model. 818

The young age of Spring Creek makes it vulnerable to hydrological extremes. Scenario simulations with the calibrated model show that without the recharge from the Wairau River, Spring Creek would run dry within approximately 300 days (dash-dot line in Figure 8). These results demonstrate that Spring Creek is very closely related to the Wairau River and any changes in the flow statistics of the river will result in a matching change at Spring Creek with little time delay.

Summary and Conclusions

In this study, we presented a model-based approach to analyse the surface water - ground 826 water exchange mechanisms in one of New Zealand's gravel-bed rivers. A highly para-827 metrized numerical model was set up for a 23 km long section of the Wairau River and 828 calibrated using different data types, regularization techniques, and the parameter es-829 timation software PEST. The trade-off between data fit and parameter homogeneity was 830 investigated with Pareto analysis methods. Null-space Monte-Carlo sampling techniques 831 were applied to estimate predictive uncertainty for data types that were used in the calib-832 ration (type I predictions) and data types that were not included in the calibration data 833 set (type II predictions). 834

Based on the results of this analysis, the following main conclusions can be drawn: 835

• The gravel aquifer underneath the river is almost exclusively recharged by river water. Land surface recharge accounts for only 1% of the long-term water balance.

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• The river is disconnected from groundwater and constantly losing over 80% of the considered river section. This causes the net exchange flows to be always positive.

• Since the stage variation in braided rivers is relatively low compared to channelized 840 rivers, the seepage rates from the disconnected sections of gravel-bed rivers also varies 841 only within a narrow range. Net exchange-flow rates have a lower threshold which 842 is typically exceeded and are capped at larger flows (~ 5 and 15 m³ s⁻¹, respectively, 843 for the Wairau River). 844

• During low-flow periods, the active channel area is reduced and river-exchange flows 845 decrease exponentially. Thus, prolonged dry periods lead to strongly reduced aquifer 846 recharge. Single flood events typically do not refill the aquifer storage due to the cap 847 on exchange flows. 848

• Shallow gravel aquifers under New Zealand's gravel-bed rivers can be extremely 849 transmissive. The mean transit time of the major spring at the Wairau Plain is 850 estimated to be less than 1 year. This makes the spring vulnerable to hydrological 851 extremes. 852

• Groundwater resources in the shallow gravel aquifer are vulnerable too. Climate 853 variation and particularly an increase in the frequency and duration of droughts will 854 cause a drastic depletion of aquifer storage. However, the system is resilient to some 855 degree, i.e. a sequence of wet years would increase aquifer storage again. 856

Previous reports on other New Zealand gravel-bed rivers suggest that the river-groundwater 857 exchange mechanisms found here for the Wairau River are typical for rivers with similar 858

settings. Local differences remain due to the specific geological settings and hydraulic
characteristics of the aquifer materials. However, the topographical setting of the of the
South Island of New Zealand has caused the formation of relatively similar low-land river
systems, particularly along the East coast. The rigour and transferability of our findings
to other gravel-bed river systems should be investigated in future studies.

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Figure 1: The Wairau Plain study site and model domain.



Figure 2: Conceptualization of the geology in the Wairau Plain model domain.



Figure 3: Historic and contemporary differential flow gaugings of the Wairau River.



Figure 4: Wairau Plain surface water - groundwater flow model.



Figure 5: Model calibration: Trade-off between measurement and regularization objective function.



Figure 6: Measured and predicted groundwater heads for the Old MCB, MCB, Catchment Board, Conders, and Pauls Rd wells. The corresponding Wairau River discharge at SH1 is shown in the top panel and 95% uncertainty bounds are shaded gray.



Figure 7: Measured and predicted groundwater heads for the P Neal, Giffords Rd, Wratts Rd, Selmes Rd, and Murphys Rd wells. The corresponding Wairau River discharge at SH1 is shown in the top panel and 95% uncertainty bounds are shaded gray.



Figure 8: Measured and predicted Spring Creek flows including the 95% predictive uncertainty bounds.



Figure 9: Boxplot of the calibration "soft targets", the corresponding simulations with the calibrated model and the 50%/95% uncertainty bounds (box/whiskers).



Figure 13: Simulated net river - groundwater exchange flow: a) the Wairau River flow, b) the net exchange flow, and c) the anomaly of net exchange flows.



Figure 10: Hydraulic conductivity fields of the calibrated model and the three different layers (left panels) and their corresponding uncertainty (right panels).



Figure 11: Specific yield fields of the calibrated model and the three different layers (left panels) and their corresponding uncertainty (right panels).



whiskers, respectively. The parameter values of the calibrated models are indicated by the blue dots. The uncertainty of Figure 12: Boxplot of normalized parameter uncertainty with the 50%/95% uncertainty bounds shown by the boxes and the K and S_y parameter fields are shown in Figures 12 and 13.



Figure 14: Snapshot of the transient model simulations for the low flow period (17/02/2014): a) the hydraulic head field, b) the spatial distribution of the drainage rates of the Wairau River and the spring/stream network of the Wairau Plain, and c) the connectivity between Wairau River and groundwater and the cell-by-cell drainage rates projected along the river length.



Figure 15: Relationship between Wairau River discharge and net river-groundwater exchange flow.



Figure 16: Cumulative travel time distribution of Spring Creek flows of the calibrated model and the corresponding uncertainty from the NSMC runs. Also shown is the cumulative distribution function of the mean travel time of these runs.

Name	# of parameters	Range
K_H (Rapaura Fm) $[m d^{-1}]$	90	$1E^{0} - 1E^{3}$
K_H (Dillons Pt. Fm) [m d ⁻¹]	1	$1E^{-1} - 5E^{1}$
S_y (Rapaura Fm) $[m^3 m^{-3}]$	90	$1E^{-4}$ - $3E^{-1}$
S_y (Dillons Pt. Fm) $[m^3 m^{-3}]$	1	$1E^{-7} - 1E^{-3}$
$S_{S} [{\rm m}^{-1}]$	4	$1E^{-7}$ - $1E^{-3}$
f_a [-]	4	$1E^{0} - 1E^{1}$
$K_R \left[\mathrm{m d^{-1}} \right]$	12	$1E^{-3}$ - $2E^{-1}$
$K_D \left[\mathrm{md^{-1}}\right]$	5	$1E^{-4} - 1E^3$

Table 1: Parameters of the Wairau Plain model and corresponding ranges used in the model calibration. Symbols are described in the text.

	Calibration Period		Evaluation Period			
Target	# of obs	RMSE	\mathbf{R}^{2}	# of obs	RMSE	\mathbf{R}^2
Old MCB	111	0.19	-0.22	284	0.45	0.12
MCB	97	0.35	0.53	284	0.55	0.36
Catchment Bd	111	0.31	0.68	284	0.21	0.89
Conders 2	925	0.24	0.84	284	0.17	0.91
Pauls Rd	199	0.30	0.84	252	0.49	0.90
P Neal	119	0.23	0.83	217	0.34	0.79
Giffords Rd	126	0.20	0.79	219	0.28	0.78
Wratts Rd	925	0.15	0.87	284	0.10	0.88
Selmes Rd	925	0.05	0.87	282	0.05	0.84
Murphys Rd	925	0.23	0.82	284	0.21	0.66
Spring Creek Flows	31	0.22	0.91	7	0.32	0.67

 Table 3: Performance of the calibrated model for groundwater head and Spring Creek flow data.

Table 2: Number of observations used	d for model calibration	n and evaluation and	their weight in the o	bjective function.
	Number of c	observations		Calibrated model
	Calibration period	Evaluation period	Observation weight	objective function
Data type	(925 days)	(284 days)	(single data point)	[-]
Groundwater head (permanent wells)	3700	1134	0.465	1 06
Groundwater head (temporary wells)	763	1540	0.465	09.4
Differential river gauging	2	0	20.0	0.4
Spring Creek flows	31	2	$1.056~\mathrm{E}^{-4}$	19.3
Mean Spring Creek flow (section 2)	1	0	$6.056~\mathrm{E}^{-5}$	$6.2~{ m E}^{-3}$
Mean flow southern streams	1	0	$1.056~\mathrm{E}^{-5}$	1.6
Total	4498	2681	I	60.7

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